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## COMPARING GROWTH AND BODY CONDITION OF INDOOR-REARED, OUTDOOR-REARED, AND DIRECT-RELEASED JUVENILE MOJAVE DESERT TORTOISES

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**Abstract.**—Desert Tortoise (*Gopherus agassizii*) populations have declined, and head-starting hatchlings in captivity until they are larger and older, and presumably more likely to survive, is one strategy being evaluated for species recovery. Previous studies have reared hatchlings in outdoor, predator-proof pens for 5–9 y before release, in efforts to produce hatchlings in excess of 100–110 mm midline carapace length that are believed to be predation-resistant. We began a comparative study to evaluate indoor-rearing to shorten this rearing period by facilitating faster initial growth. We assigned 70 neonates from the 2015 hatching season to three treatment groups: (1) indoor-reared (n = 30), (2) outdoor-reared (n = 20), and (3) direct-release (n = 20). We released direct-release hatchlings shortly after hatching in September 2015 and monitored them 1–2 times per week with radio telemetry. We head-started the indoor- and outdoor-reared treatment groups for 7 mo before releasing them in April 2016. Indoor-reared tortoises were fed five times per week (September to March). Outdoor-reared tortoises had access to native forage and we gave them supplemental water and food once per week while active before winter dormancy. Indoor-reared tortoises grew > 16 times faster than direct-release tortoises and > 8 times faster than outdoor-reared tortoises; however, indoor-reared tortoises weighed less and had softer shells than comparatively sized older (3–4 y-old) tortoises raised outdoors. Increasing the duration of the indoor-rearing period or incorporating a combination of both indoor and later outdoor husbandry may increase shell hardness among head-starts, while retaining the growth-promoting effect of indoor rearing and shortening overall captivity duration.

**Key Words.**—body condition; Chelonian; conservation; husbandry; morphology; reptile; threatened; wildlife management

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### INTRODUCTION

Head-starting seeks to increase the number of animals eventually recruited into a breeding population by raising juvenile animals in protected conditions early in life and releasing them into the natural environment at a larger size when they are presumably more likely to survive (Heppell et al. 1996; Burke 2015). Head-starting projects have been initiated with varying success for mammals (Sinn et al. 2014), birds (Cohn 1999), amphibians (Lannoo 2005), and reptiles (Jarvie et al. 2015; Tuberville et al. 2015). Turtles may be particularly suited to head-starting as a recovery tool (Burke 2015) because they have low survival in the wild during their early life stages and high survival as adults under most natural conditions (Gibbons 1987). Turtle head-starting studies have increased recently (see *Herpetological Conservation and Biology*, Volume 10), and include Blanding's Turtles (*Emydoidea blandingii*;

Green 2015; Buhlmann et al. 2015), Gopher Tortoises (*Gopherus polyphemus*; Tuberville et al. 2015; Quinn et al. 2018), Western Pond Turtles (*Actinemys marmorata*; Vander Haegen et al. 2009), and Kemp's Ridley Sea Turtles (*Lepidochelys kempii*; Caillouet et al. 2015), among others.

Head-starting can be a useful tool in turtle conservation. For example, head-starting has been used to reestablish wild populations of Blanding's Turtles (Buhlmann et al. 2015) and Galapagos Tortoises (*Chelonoidis hoodensis*; Gibbs et al. 2014) in areas where they had previously been extirpated. Head-starting has also been useful in restoring ecosystem services. By establishing populations of the non-native Aldabra Giant Tortoise (*Aldabrachelys gigantea*) using head-starting to replace extinct *Cylindraspis*, conservationists have begun to control the spread of invasive alien species, as the Aldabra tortoises restored grazing and seed dispersal to the ecosystem (Griffiths et

al. 2010; Vikash et al. 2018). Modeling can indicate how populations are most likely to respond to head-starting. In some scenarios (e.g., Spencer et al. 2017), head-starting can lead to successful conservation outcomes even if underlying threats cannot be abated; however, this is not generally the case (Heppell et al. 1996; Reed et al. 2009). Although modeling can be useful, long-term, post-release monitoring is necessary to fully evaluate the efficacy of head-starting (Buhlmann et al. 2015; Burke 2015; Nagy et al. 2015b).

Mojave Desert Tortoise (*Gopherus agassizii*) populations have declined throughout their range (Berry 1986; U.S. Fish and Wildlife Service [USFWS] 1990, 2011). Habitat loss, increased mortality from roads (i.e., automobiles), human-subsidized predators (e.g., Common Ravens, *Corvus corax*, and Coyotes, *Canis latrans*), upper-respiratory tract disease (*Mycoplasma* spp.), and habitat degradation from disturbance and invasive plants have all been identified as contributing causes (Berry 1986; Esque et al. 2010; Nafus et al. 2013; Peaden et al. 2015). Head-starting has been identified as a possible management action to reinforce diminished populations of Mojave Desert Tortoises (USFWS 2008, 2011; hereafter desert tortoises), provided that the original causes of population decline have been mitigated or are addressed concurrently. Several desert tortoise head-starting facilities have begun evaluating the efficacy of rearing hatchling tortoises in predator-proof outdoor pens before releasing them into the wild (Hazard and Morafka 2002; Nafus et al. 2015; Nagy et al. 2015b). Estimates of size at which juvenile post-release survival substantially increases range from 84 mm (Hazard et al. 2015) to 100 mm mid-line carapace length (MCL; Nagy et al. 2015b). Although supplemental food and water can increase growth and survival of desert tortoises raised outdoors (Nafus et al. 2017, Nagy et al. 2015a), outdoor rearing, as in the wild, may take 5–9 y to produce a juvenile tortoise of 100 mm MCL (Nagy et al. 2015a) because tortoises maintain natural behaviors and are inactive during both the hottest and coldest seasons in the desert. Rearing tortoises indoors may decrease the time needed to raise tortoises to larger size by keeping juveniles active and growing during the winter months, when growth otherwise ceases in the wild. No study has yet evaluated indoor head-starting in desert tortoises.

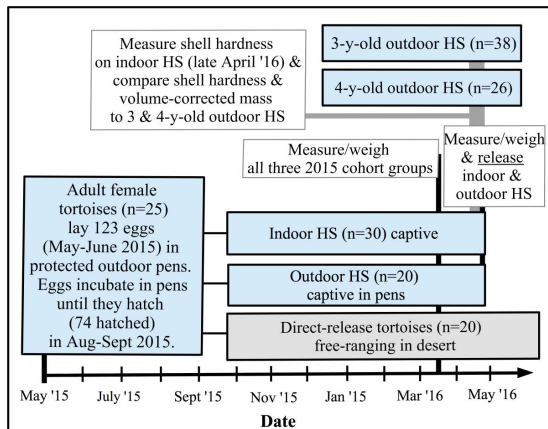
Accelerating growth of captive tortoises by raising them indoors may have unknown consequences for the overall health and condition of the animals. Thus, the monitoring of metrics like body condition and shell hardness within head-start studies would provide a comprehensive assessment of robustness of captive-reared tortoises. Body condition (BC) is expressed as the mass of an animal relative to an appropriate

size metric (approximated volume in our case) and can reflect nutritional condition, stored fat, and water balance (Shine et al. 2001; Nagy et al. 2002; Loehr et al. 2007; Nagy et al. 2015a). Shell hardness increases with body size and age in juvenile desert tortoises (Nagy et al. 2011), and the hardness of the shell of a turtle likely plays a major role in its protection against predators.

The goal of our study was to evaluate the feasibility of indoor rearing to reduce the time needed to head-start desert tortoises relative to outdoor rearing. The research presented here is part of a larger effort to evaluate indoor head-starting through long-term, post-release monitoring. We established three treatment groups: indoor-reared head-started tortoises, outdoor-reared head-started tortoises, and direct-release hatchlings (all 2015 cohort). We reared indoor and outdoor head-start animals for seven months (September to April) prior to release, and direct-release animals were released in the natural environment days after hatching (September) to serve as a control. We compared growth, body condition, and survival among the three treatment groups at the end of the 7-mo period (a short time relative to the potential lifespan of > 50 y of desert tortoises). We also evaluated the indoor-reared tortoise group for shell hardness at the end of the rearing period relative to similarly sized, but older (3–4 y-old of 2011 and 2012 cohorts) outdoor-reared captive tortoises from an earlier study.

## MATERIALS AND METHODS

**Study site.**—The Mojave National Preserve (MNP) is a 650,000 ha preserve in San Bernardino County, California, USA, in the eastern Mojave Desert managed by the U.S. National Park Service (NPS). We conducted all experiments and observations in Ivanpah Valley in the northeastern part of the Mojave National Preserve. The primary habitat in Ivanpah Valley is Creosote Bush Scrub and is dominated by Creosote Bush (*Larrea tridentata*), White Bursage (*Ambrosia dumosa*), and low-density *Yucca* (*Y. schidigera*, and *Y. brevifolia*; Turner et al. 1984, Todd et al. 2016). Although tortoises are commonly seen in Ivanpah Valley and habitat suitability is relatively high (Nussear et al. 2009), current tortoise densities are much lower than they were historically (3.8 tortoises per km<sup>2</sup> in 2008, Allison 2012; 77–85 tortoises per km<sup>2</sup> in 1977–1980, Turner et al. 1984). The area into which we released juvenile tortoises was a 0.7-km<sup>2</sup> unfenced plot centered 850 m from a powerline service road in the Ivanpah Valley of the MNP, the exact location not disclosed for security (Lindenmayer and Scheele 2017; Litzgus, J. 2017. The illegal turtle trade: why scientists keep secrets. Available from <http://theconversation.com/the-illegal-turtle-trade-why-i-keep-secrets-85805>. [Accessed 29 November



**FIGURE 1.** Diagram showing study design for experiment to test effects of indoor-rearing on Mojave Desert Tortoise (*Gopherus agassizii*) growth, body condition (volume-corrected mass), and shell hardness. This study is part of a larger effort to evaluate the effectiveness of indoor-rearing in head-starting desert tortoises. Blue-filled boxes represent groups of tortoises (or tortoise eggs) held temporarily in captivity, whereas the grey-filled box represents released, free-ranging tortoises (a control group). White boxes indicate analyses of morphological data from different times in the study.

2018]). Other human disturbances in our study area include abandoned cattle grazing infrastructure (fencing and corrals), and a railroad track 4.5 km away.

**Obtaining hatchlings.**—In May 2015 we captured our previously radio-tagged female desert tortoises (Nafus et al. 2015), brought them to the Ivanpah Desert Tortoise Research Facility (hereafter, facility), and radiographed them to determine gravidity. We placed females with at least three calcified eggs in predator-proof nesting pens at the facility; we returned all others to their capture sites. The nesting enclosure measured 30 × 30 m and was permanently subdivided with metal siding into 18 smaller pens (5 × 9 m). We constructed artificial burrows for each female tortoise to use as shelter and nesting sites. Burrows were at least 1 m in length and were constructed from 310-mm (12-in) diameter cardboard Quik-Tube building forms® (Quikrete International, Atlanta, Georgia, USA) that were cut in half longitudinally and buried at a 20° angle underground. We kept gravid females until they laid their eggs (within 30 d in most cases) and then returned them to their last burrow location. As hatching approached (70 d into the estimated 90-d incubation period), we began monitoring pens for hatchlings several times daily.

We obtained 74 hatchlings and temporarily housed them by clutch until all hatchlings had emerged. We permanently marked each neonate by using nail clippers to notch the marginal scutes with a unique identification pattern (modified from Cagle 1939) with codes assigned

to us by USFWS. We excluded four hatchlings from the study due to especially low body mass at hatching and/or developmental defects. We assigned healthy hatchlings to one of three treatment groups (described further below): (1) direct-release (control group,  $n = 20$ ); (2) outdoor-reared ( $n = 20$ ), and (3) indoor-reared ( $n = 30$ ; Fig. 1). Within each clutch, we randomly assigned individuals to treatment groups, attempting to divide each clutch as evenly as possible among treatment groups to avoid confounding clutch and treatment effects. When more than one hatchling from a clutch was assigned to a treatment, we distributed clutch-mates among replicate enclosures receiving the same treatment.

**Experimental treatment groups.**—On 24 September 2015 (21–46 d after hatching), we moved the indoor-reared head-start (HS) treatment group to mesocosms inside the climate-controlled facility. We set ambient temperature inside the facility to a constant 24.4° C. We used 189-L (50-gallon) Rubbermaid (Atlanta, Georgia, USA) stock tanks (132 × 79 × 30.5 cm) filled with a 5-cm layer of natural desert sand as substrate. We established six tanks, each of which housed five tortoises. We suspended Mini Combo Deep Dome Dual Lamp Fixtures (ZooMed® Laboratories Inc., San Luis Obispo, California, USA) over tanks and each held a 50-Watt ZooMed® Repti Basking Spot Lamp bulb for daytime basking and a ZooMed® 50-Watt Infrared Basking Spot bulb on the other side for nighttime heat. Lights were connected to automatic timers. We timed basking lights to operate 0600–1830, and the infrared lights were timed to operate 1900–0530 (an interval set to approximate the natural photoperiod at the beginning of the study, but not adjusted seasonally thereafter). The lights created basking spots of 37° C during the day and 32° C at night. The daytime basking bulbs provided ultraviolet long-wave light (UVA) but not ultraviolet short-wave (UVB) light. Windows on two sides of the room also provided some natural light during the day.

We outfitted each tank with three cover items constructed from halved plastic pipe (11.5 cm in diameter and cut into 12-cm linear segments) and a paper feeding plate. Because inadequate humidity has been linked to unnatural shell growth as tortoises grow (Wiesner and Iben 2003), we provided a humid hide box in each mesocosm to promote smooth shell growth. We maintained humidity in the hide boxes by cutting burrow-shaped entrance holes (one hole per hide box) into the sides of lidded plastic tote boxes (Rubbermaid Roughnecks; Rubbermaid, Atlanta, Georgia, USA; 40 × 26 × 18 cm) and we lined each tote box with 7 cm of moist peat moss that was re-moistened every 3–4 d.

We fed indoor-reared hatchlings ad libitum five times per week and soaked them weekly for 15 min in 1–2 cm

of water to allow them to drink. Diet was a mixture, by mass, of leafy greens (50%) supplemented with commercially available food pellets (25%, ZooMed® Grassland Tortoise Diet, ZooMed® Laboratories Inc., San Luis Obispo, California, USA), and 25% water (used to soften the pellets). At each feeding, the greens mixture consisted of equal amounts of five leafy greens readily available at grocery stores and selected to closely approximate the nutritional properties of the natural forage of desert tortoises. Collectively, they provide the ratio of phosphorus to potassium that facilitates assimilation of calcium (Jarchow et al. 2002). The greens included Dandelion (*Taraxacum officinale*), Mustard greens (*Brassica juncea*), Turnip greens (*Brassica rapa* var. *rapa*), Collards (a cultivar of *Brassica oleracea*), and Endive (*Cichorium endivia*). If one of the preferred choices was unavailable locally, we used Kale (a cultivar of *Brassica oleracea*) or Swiss Chard (*Beta vulgaris cicla*) as a substitute. On 11 December 2015, we also began adding Rep-Cal® Calcium with Vitamin D<sub>3</sub> (Rep-Cal Research Labs, Los Gatos, California, USA) to the food mixture twice per week.

On 23 September 2015, we placed outdoor-reared hatchlings into predator-proof, semi-natural pens at the facility. The 30 × 30-m enclosure was constructed of chain-link fence (buried to exclude digging mammals) and was covered with netting to exclude avian predators. Within the larger enclosure, we placed hatchlings into two 10 × 10-m pens at a density of 10 hatchlings per pen. The pens mimicked the local natural environment and contained native vegetation, sand substrate, rocks, dead woody structure, and starter burrows. The starter burrows were constructed from halved PVC pipe (13-cm diameter) buried in a 0.5–1.0-m trench at a 20° angle from the surface.

We provided artificial rain weekly during the active season (until late October) for 30 min with rotating sprinklers to allow hatchlings to drink and to stimulate growth of native vegetation (Beatley 1974). Within the outdoor rearing pens, we inventoried 11 annual species on which tortoises are known to forage (Jennings and Berry, 2015; Abella and Berry, 2016) including, Wingnut Cryptantha (*Cryptantha pterocarya*), Common Stork's-bill (*Erodium cicutarium*), Common Mediterranean Grass (*Schismus barbatus*), Desert Indianwheat (*Plantago ovata*), Booth's Evening Primrose (*Camissonia boothii*), Brittle Spineflower (*Chorizanthe brevicornu*), Devil's Spineflower (*C. rigida*), Pepperweed (*Lepidium* spp.), Desertsnow (*Linanthus demissus*), Desert Dandelion (*Malacothrix glabrata*), and Whitestem Blazingstar (*Mentzelia albicaulis*). Although natural vegetation was readily available for outdoor-reared tortoises, we also provided supplemental food on watering days because watering stimulated hatchling exploration and feeding (Nafus et

al. 2017). The food mixture was the same as described above for the indoor-reared hatchlings. While tortoises were housed within pens, we frequently observed foraging behavior on both natural vegetation and supplemental food; however, we collected no data on their foraging preferences. The amount of supplemental food fed was 5% of the total tortoise biomass in each pen, which was functionally ad libitum, but which minimized waste to avoid attracting ants (Formicidae). We ceased supplemental watering and feeding during the fall and winter when hatchlings are normally inactive outdoors.

On 28 September 2015, we released hatchlings assigned to the direct-release treatment group (hereafter DR) into the natural environment in Ivanpah Valley. We released the hatchlings within a 0.7-km<sup>2</sup> rectangular study area of unmanipulated and unfenced natural desert that was imbedded in the general area where their mothers had been captured. We used radio-telemetry to monitor these free-ranging hatchlings for post-release growth and survivorship. We attached 2.1-g radio transmitters (BD-2, Holohil Systems Ltd., Ontario, Canada) with 7-mo batteries to each tortoise on the fourth vertebral scute with 5-min epoxy. The mass of radio transmitters was generally < 10% of the mean body mass for released hatchlings (range = 7.7–11.9%, with only 3/20, or 15%, of individuals with transmitters > 10% of their body mass). Prior to release, we monitored animals for any signs of stress or abnormal behavior. Radio transmitter placement and type were approved prior to release by USFWS and California Department of Fish and Wildlife. We used a 3-element Yagi antenna (AF Antronics, Inc., Urbana, Illinois, USA) and a R1000 receiver (Communications Specialists, Inc., Orange, California, USA) to locate each animal daily for the first 4 d following release, and then twice per week through the duration of the active season (until 12 November 2015). Throughout the winter, tracking frequency was reduced to once per week. We resumed twice per week tracking in March 2016.

**Morphometrics.**—We measured and weighed (hereafter measured) hatchlings immediately after hatching (hatching size) and then again prior to treatment group assignment on 22 September 2015 (initial size). We measured indoor-reared tortoises approximately every 30 d until their release in late April 2016. We measured DR tortoises again during 11–15 March 2016 when we replaced their radio transmitters. We measured indoor and outdoor HS animals on 16 March 2016 to facilitate comparison among the three treatment groups. We measured the indoor and outdoor HS animals again prior to their release in late April 2016. We recorded mass with a digital scale to the nearest 0.01 g. We measured the following to the nearest 0.1 mm using vernier calipers: (1) midline carapace length



(MCL, straight-line distance from the anterior edge of the nuchal scute to the inside of the natural notch in the supracaudal scute), (2) maximum shell height, and (3) maximum shell width on the bridge.

**Body condition.**—We calculated body condition (BC) for all surviving animals from all treatment groups based on measurements taken 5.8 mo into the rearing period (11–16 March 2016) using the formula described by Loehr et al. (2004), where  $BC = \text{body mass (g)} / \text{shell volume (cm}^3\text{)}$ . We computed shell volume (for the BC calculations and for analysis of volume-corrected mass) using the standard formula for a half-ellipsoid (Loehr et al. 2004), where all input sizes are in mm and the product is in  $\text{cm}^3$ :

$$\text{Shell volume (cm}^3\text{)} = (\pi \times \text{MCL} \times \text{width} \times \text{height}) / 6000$$

To make our data easily comparable with metrics most often reported in the literature, we also present body condition in the formula described by Nagy et al. (2002), where shell volume, approximated as a box, is calculated as  $\text{MCL} \times \text{width} \times \text{height}$  (Tables 1–2).

**Shell hardness.**—To measure shell hardness, we used a 10.2-cm (4-in) tension-calibrated micrometer (model 3732XFL-4; L.S. Starrett Company, Athol, Massachusetts, USA) to measure normal, uncompressed shell height (UCSH) at the center of the third vertebral scute (Nagy et al. 2011). We then turned the micrometer spindle, compressing the shell of each tortoise between the two measuring faces until a point where the spindle ratchet began to slip continually for approximately  $240^\circ$  of further turning. We then read the micrometer for a compressed shell height (CSH) reading. We calculated shell hardness index (SHI) as described by Nagy et al. (2011), where an index value of 100 corresponds to complete hardness:

$$\text{Shell Hardness Index (SHI)} = (\text{CSH} / \text{UCSH}) \times 100$$

We measured shell hardness of surviving indoor HS animals ( $n = 29$ ) just prior to release in April 2016. Prior to analysis, we discarded an unrealistically low measurement attributed to one indoor HS individual, which we suspect was due to misreading the instrument. Of the 2015 cohort juveniles, we were only able to measure the indoor HS group for shell hardness. The 2015 cohort outdoor HS and DR juveniles were too soft to safely measure, raising concern that compressing these smaller tortoises would cause injury. Therefore, we compared the shell hardness of the 2015 cohort indoor HS juveniles with shell hardness data taken in September 2015 from similar-sized, but older (3–4

y-old), outdoor-reared animals (2011–2012 cohorts) from other enclosures at the facility (Nafus et al. 2017). Similarly, we also compared body condition among indoor HS 2015 cohort juveniles and similar-sized but older 2011 and 2012 cohort juveniles in case any differences observed among the 2015 treatment groups (indoor HS, outdoor HS, and DR) were a result of allometric effects. These 2011 and 2012 cohort juveniles were reared in similar conditions to the 2015 cohort, outdoor HS animals in our current study, except approximately half of them (19/38: 2012 cohort; 14/26: 2011 cohort) were rain supplemented half as often (one time per two weeks vs. one time per one week), as part of a previous investigation on the effects of differing artificial rainfall regimens on head-starting (Nafus et al. 2017).

**Statistical methods.**—We performed all statistical tests in Program R (R Core Team 2014), using  $\alpha = 0.05$  as the acceptable threshold of Type I error. We present data as means  $\pm$  1 SE. We used the Kaplan-Meier estimator to estimate the survival of each treatment group (Pollock et al. 1989). To test for MCL differences among treatment groups (2015 cohort), we used linear mixed effects models with MCL as the response variable and the unique ID code of the mother as a random effect. Visual assessment of histograms of model residuals showed that model residuals generally approximated the normal distribution, with the exception of one outlier (the smallest individual in the indoor-reared treatment group). We ran our analyses with and without this outlier, and outcomes were unchanged; thus, we felt justified modeling our full dataset (including the outlier).

To analyze volume-corrected mass (a way of analyzing body condition) by treatment group (2015 cohort), we used analysis of covariance (ANCOVA) with log-transformed mass as the response variable and log-transformed shell volume as the covariate (García-Berthou, 2001), and we included maternal identity as a random effect. We also used ANCOVA to test for treatment group (indoor HS 2015 cohort, outdoor HS 2011 cohort, and outdoor HS 2012 cohort) effects on shell hardness and volume-corrected mass. When analyzing shell hardness, we included MCL as a covariate, and when analyzing volume-corrected mass we used log-transformed mass as the response variable and log-transformed shell volume as the covariate (García-Berthou, 2001). For each ANCOVA, we tested for interactions between treatment variables and covariates and retained interactive terms when significant. When treatment effects were detected, we performed Tukey's post-hoc multiple comparisons using the `glht` function in the `multcomp` package in R to further investigate treatment group differences.

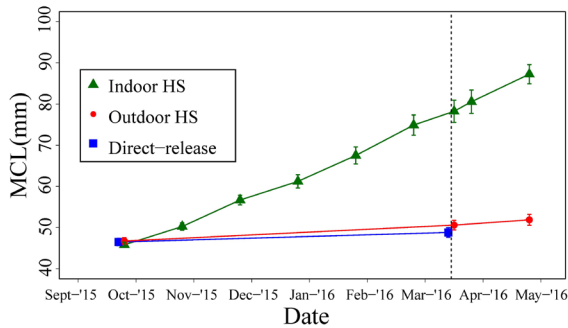


FIGURE 2. Means  $\pm$  95% confidence intervals for midline carapace length (MCL) of 2015 cohort juvenile Mojave Desert Tortoises (*Gopherus agassizii*) under three husbandry treatments. The dashed vertical line indicates time when data were collected for comparative analyses among the three treatment groups.

### RESULTS

Twenty-five of 31 captured females were gravid with at least three eggs and were placed in predator-proof nesting pens at the facility. They collectively laid 123 eggs, from which 74 hatchlings successfully emerged (60.2% emergence success). All (20/20) outdoor HS tortoises survived until their release on 25 April 2016. All indoor HS tortoises (30/30) survived in captivity until mid-March, when data were taken for comparative purposes. On 8 April 2016, however, we found one indoor HS dead of unknown causes in its mesocosm; thus, survival for the indoor HS group was 96.7% (29/30) through the head-start period. Fifteen of the 20 (75%) DR juveniles survived from their release on 28 September 2015 until spring measurements in mid-March 2016; we found four dead and we never found one (unknown fate) before mid-March. Throughout the rearing period, survival estimates among the treatment groups did not differ significantly ( $P > 0.05$  based on overlapping confidence intervals; Daly 2017).

Mean initial MCL for the 2015 cohort was  $46.9 \pm 0.2$  mm ( $n = 70$ ; range, 40.5–50.2) and MCL did not differ significantly ( $F_{2,52} = 2.67, P = 0.078$ ) among the three treatment groups (DR:  $46.5 \pm 0.4$  mm; indoor HS:  $45.8 \pm 0.3$  mm; outdoor HS:  $46.8 \pm 0.4$  mm). After 5.8 mo (mid-March 2016), mean MCL differed significantly ( $F_{2,47} = 249.3, P < 0.001$ ; Fig. 2; Table 1) among treatment groups (DR:  $48.8 \pm 1.4$  mm; indoor HS:  $78.2 \pm 1.0$  mm; outdoor HS:  $50.6 \pm 1.2$  mm). Indoor HS tortoises were larger (MCL) than either outdoor HS ( $z = 18.67, P < 0.001$ ) or DR juveniles ( $z = 18.16, P < 0.001$ ), whereas outdoor HS and DR juveniles did not differ significantly in MCL ( $z = 1.05, P = 0.543$ ). Among the 2015 cohort, indoor HS tortoises grew over 16 times faster in length than DR tortoises ( $70.7 \pm 2.0\%$  vs.  $4.3 \pm 2.7\%$ ) and over eight times faster during the first 5.8 mo than outdoor HS tortoises ( $70.7 \pm 2.0\%$  vs.  $8.3 \pm 2.4\%$ ).

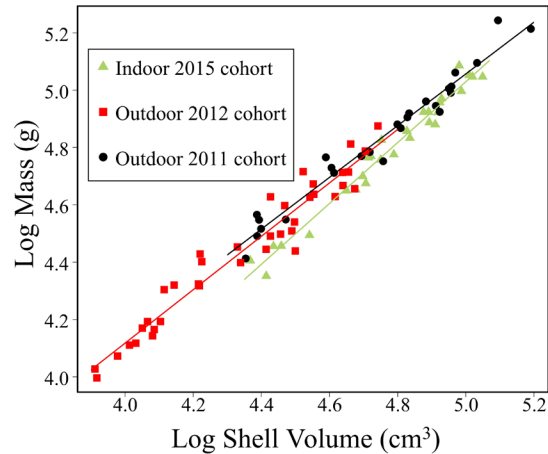
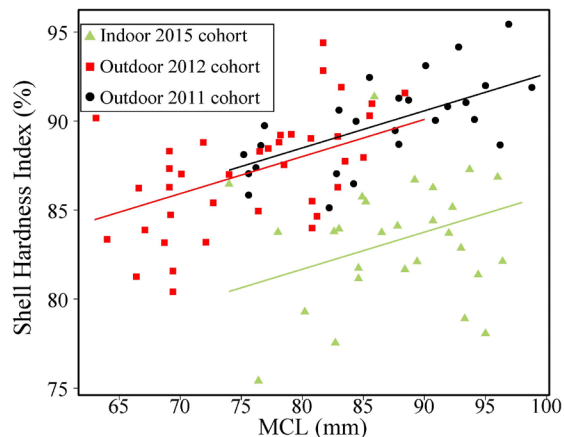


FIGURE 3. Log-scaled body mass (grams) versus log-scaled shell volume ( $\text{cm}^3$ ) for three treatment groups of juvenile Mojave Desert Tortoises (*Gopherus agassizii*): (1) indoor-reared 2015-cohort juveniles:  $n = 29$ ; age = 7.5 mo, (2) outdoor-reared 2012-cohort juveniles:  $n = 38$ , age = 3 y, and (3) outdoor-reared 2011-cohort juveniles:  $n = 26$ , age = 4 y.

Mean initial mass was  $22.6 \pm 0.3$  g ( $n = 70$ ) and mass did not differ significantly ( $F_{2,51} = 1.95, P = 0.153$ ) among treatment groups (DR:  $22.6 \pm 0.6$  g; indoor HS:  $22.3 \pm 0.5$  g; outdoor HS:  $23.1 \pm 0.8$  g). After 5.8 mo (mid-March 2016) the indoor HS were heavier than the outdoor HS or DR tortoises (DR:  $24.5 \pm 3.7$  g; indoor HS:  $94.7 \pm 2.7$  g; outdoor HS:  $30.2 \pm 3.3$  g); however, after taking into account shell volume, relative mass did not differ significantly among the treatment groups (indoor HS vs. DR:  $t = -0.387, df = 46, P = 0.913$ ; indoor HS vs. outdoor HS:  $t = -1.198, df = 46, P = 0.432$ ; outdoor HS vs. DR:  $t = 1.676, df = 46, P = 0.202$ ). Mean ratio values for body condition (box and half-ellipsoid type) from mid-March 2016 were similar among treatment groups (Table 1). When comparing volume-corrected mass of 2015 indoor HS tortoises with older, but similar-sized (2011 and 2012 cohort) outdoor-reared tortoises from a previous study, we found a significant difference among treatment groups ( $F_{2,87} = 660.83, P < 0.001$ ) and a significant interaction between treatment group and log-transformed shell volume ( $F_{2,87} = 3.66, P = 0.030$ ; Fig. 3). Among all smaller-size tortoises, volume-adjusted mass of indoor HS tortoises was less than that of comparatively sized but older outdoor HS tortoises (Fig. 3); however, this difference diminished as tortoises approached 95 mm in MCL (Fig. 3).

Midline carapace length was a significant predictor of shell hardness ( $\beta = 0.208, P < 0.001$ ). For every 1.0-mm increase in MCL, shell hardness increased by 0.21% (Fig. 4), and there was no evidence ( $F_{2,86} = 1.66, P = 0.196$ ) that this relationship differed by cohort. Indoor HS tortoises (2015 cohort) had a mean shell hardness index of  $83.2 \pm 0.6\%$ . Mean shell hardness index among



**FIGURE 4.** Shell hardness index (% of total hardness) versus midline carapace length (MCL) for three treatment groups of juvenile Mojave Desert Tortoises (*Gopherus agassizii*): (1) indoor-reared 2015-cohort juveniles:  $n = 28$ ; age = 7.5 mo, (2) outdoor-reared 2012-cohort juveniles:  $n = 38$ , age = 3 y, and (3) outdoor-reared 2011-cohort juveniles:  $n = 26$ , age = 4 y.

older outdoor HS tortoises was  $89.9 \pm 0.6\%$  for the 2011 cohort ( $n = 26$ ) and  $87.1 \pm 0.5\%$  for the 2012 cohort ( $n = 38$ ; Fig. 4; Table 2). At a fixed size of 80 mm in MCL, outdoor-reared 2012 and 2011 cohort tortoises both had predicted shell hardness index of 88%, whereas the 2015 cohort indoor-reared tortoises had a predicted shell hardness index of 82% (Fig. 4). There was a significant

difference in shell hardness among the groups ( $F_{2,88} = 38.32$ ,  $P < 0.001$ ). With MCL included as a covariate, indoor HS tortoises had significantly softer shells (lower shell hardness index) than did either 2011- ( $t = -8.93$ ,  $df = 88$ ,  $P < 0.001$ ) or 2012-cohort ( $t = -7.32$ ,  $df = 88$ ,  $P < 0.001$ ) outdoor HS animals. In other words, a 2015 indoor HS tortoise (at 7.5 mo age) was likely to have a softer shell than either a 2011- or 2012-cohort, outdoor HS tortoise of the same size. However, outdoor-reared 2011- and 2012-cohort HS tortoises did not differ from one another in shell hardness ( $t = 0.57$ ,  $df = 88$ ,  $P = 0.837$ ; Fig. 4; Table 2).

## DISCUSSION

Over their first 7 mo of life, tortoises reared indoors grew much faster than siblings from the same cohort either released immediately into the wild after hatching (DR) or reared outdoors in protected enclosures. Survival during indoor and outdoor head-starting was high, 97% and 100%, respectively, compared to 75% survival of direct-release hatchlings over the same period. After 7 mo, indoor-reared tortoises reached the same mean size (87 mm in MCL) as 5–6-y-old wild desert tortoises (Turner et al. 1987; Curtin et al. 2009). Additionally, indoor-reared tortoises reached and exceeded, in some cases, MCL of both 2011 and 2012 outdoor-reared

**TABLE 1.** Summary statistics of growth metrics for juvenile (2015 cohort) Mojave Desert Tortoises (*Gopherus agassizii*) reared under three different husbandry treatments (direct-release, indoor-reared, outdoor-reared). Measurements were taken in mid-March 2016 ('final', 5.8 mo into the 7-mo rearing period). Intervals are reported at 95% confidence. P-values correspond to linear mixed effects models (with mothers' identities as random effect). Abbreviations are  $n$  = sample size, SE = standard error, LCI = lower 95% confidence limit, UCI = upper 95% confidence limit, Min = minimum, Max = maximum, and MCL = midline carapace length.

Metric	Treatment	n	Mean	SE	LCI	UCI	Min	Max	P-value
Final MCL (mm)	Direct-release	15	48.8	1.4	46.1	51.5	44.7	52.8	< 0.001
	Indoor	30	78.2	1.0	76.2	80.1	54.6	87.1	
	Outdoor	20	50.6	1.2	48.3	53.0	45.4	55.4	
MCL growth (%)	Direct-release	15	4.3	2.7	1.4	9.6	0.4	7.5	< 0.001
	Indoor	30	70.7	2.0	66.7	74.9	17.4	89.6	
	Outdoor	20	8.3	2.4	3.8	13.5	3.8	16.6	
Final mass (g)	Direct-release	15	24.5	3.7	16.8	31.6	17.2	30.4	< 0.001
	Indoor	30	94.7	2.7	89.1	100.0	41.0	119.4	
	Outdoor	20	30.2	3.3	23.7	36.8	20.5	39.0	
Mass growth (%)	Direct-release	15	8.8	6.6	-4.7	21.8	0.9	16.1	< 0.001
	Indoor	30	147.8	5.1	138.7	159.1	33.9	209.7	
	Outdoor	20	17.3	6.0	7.9	31.9	7.0	34.4	
Body condition (half-ellipsoid) (g/cm <sup>3</sup> ) (Loehr et al. 2004)	Direct-release	15	1.09	0.01	1.04	1.13	0.92	1.22	–
	Indoor	30	1.06	0.01	1.04	1.07	0.97	1.15	
	Outdoor	20	1.12	0.01	1.10	1.14	1.03	1.21	
Body condition (box) (g/cm <sup>3</sup> ) (Nagy et al. 2011)	Direct-release	15	0.570	0.008	0.547	0.594	0.482	0.639	–
	Indoor	30	0.554	0.006	0.547	0.562	0.510	0.602	
	Outdoor	20	0.586	0.007	0.577	0.596	0.537	0.631	

**TABLE 2.** Summary statistics of growth metrics from measurements of juvenile Mojave Desert Tortoises (*Gopherus agassizii*) taken prior to head-start release in late April 2016. Shell hardness and body condition data from indoor-reared head-starts (2015 cohort) are compared with data from outdoor-reared head-starts from 2012- and 2011-cohorts taken in September 2015. Intervals are reported at 95% confidence. Abbreviations are n = sample size, SE = standard error, LCI = lower 95% confidence limit, UCI = upper 95% confidence limit, Min = minimum, Max = maximum, and MCL = midline carapace length.

Metric	Treatment	n	Mean	SE	Lower CI	Upper CI	Min	Max
Comparison of 2015 indoor- and outdoor-reared head-starts								
Final MCL (mm)	Indoor	29	87.2	1.0	85.2	89.2	74	96.4
	Outdoor	20	51.9	1.6	49.6	54.2	46.7	58.2
Final mass (g)	Indoor	29	122.6	3.7	115.1	130.0	77.6	161.7
	Outdoor	20	30.9	4.5	22.3	39.8	22.4	42.5
Final body cond. (half-ellipsoid) (g/cm <sup>3</sup> ) (Loehr et al. 2007)	Indoor	29	1.01	0.01	1.00	1.03	0.94	1.11
	Outdoor	20	1.04	0.01	1.03	1.06	0.97	1.17
Final body cond. (box) (g/cm <sup>3</sup> ) (Nagy et al. 2011)	Indoor	29	0.531	0.005	0.525	0.539	0.492	0.582
	Outdoor	20	0.546	0.005	0.537	0.555	0.506	0.612
Comparison of 2015 indoor-reared group to outdoor-reared 2012 and 2011 cohorts								
Shell hardness index (SHI) (% of total hardness)	Indoor 2015	28	83.2	0.6	81.0	83.3	75.4	91.4
	Outdoor 2012	38	87.1	0.5	87.4	89.6	80.4	94.4
	Outdoor 2011	26	89.9	0.6	87.8	90.1	85.1	95.4
Body condition (half-ellipsoid) (g/cm <sup>3</sup> ) (Loehr et al. 2007)	Indoor 2015	28	1.01	0.01	1.00	1.02	0.94	1.11
	Outdoor 2012	38	1.08	0.01	1.06	1.10	0.94	1.23
	Outdoor 2011	26	1.10	0.01	1.08	1.12	0.99	1.20
Body condition (box) (g/cm <sup>3</sup> ) (Nagy et al. 2011)	Indoor 2015	28	0.527	0.007	0.512	0.542	0.492	0.582
	Outdoor 2012	38	0.566	0.012	0.544	0.588	0.493	0.645
	Outdoor 2011	26	0.575	0.011	0.554	0.596	0.521	0.626

tortoises, which had ranges of 75–99 mm (2011) and 63–88 mm (2012). At this rapid rate of growth (4.3 mm/mo), indoor-reared tortoises would have reached the 100 mm MCL release-size threshold recommended by Nagy et al. (2015b) in just three more months (by July 2016; a hot month, and not ideal for release). However, by September 2016, when temperatures would again be favorable for release, they would have been 109 mm in average MCL.

Of the 2015 cohort animals, only the indoor head-starts had shells hard enough to perform shell hardness measurements at 7.5 mo of age. Thus, we compared shell hardness of 2015 indoor head-starts to that of older 2011 and 2012 outdoor-reared tortoises. The younger indoor-reared tortoises had significantly lower shell hardness values than the older cohorts, but this finding is not entirely unexpected. Younger juveniles have less ossified shells than adults and desert tortoise shells become harder with age (Boarman 2003; Nagy et al. 2011). Our indoor-reared tortoises had shell hardness values (SHI = 83%) similar to 1-y-old 40-mm MCL outdoor-reared tortoises reported by Nagy et al. (2011),

but not surprisingly were softer than similar-sized, but older, outdoor reared tortoises (87.1%, 3 y-old; 89.9%, 4 y-old). Collectively, our findings and those of Nagy et al. (2011) suggest that age, not just size, plays an important role in juvenile shell hardness.

Analysis of volume-corrected mass (a way of analyzing body condition) showed that there were no differences among the 2015 cohort, and ratio values of body condition of all treatment groups fell within the range expected for healthy animals. Based on the Loehr et al. (2004) formula, mean body condition of the three treatment groups ranged from 1.06–1.12 g/cm<sup>3</sup>, with both the lowest (0.90 g/cm<sup>3</sup>) and highest values (1.22 g/cm<sup>3</sup>) across all treatments coming from direct-release animals. Using the same formula, body condition of wild, free-ranging juvenile Namaqualand Speckled Tortoises (*Homopus signatus signatus*) ranged from 0.99–1.11 g/cm<sup>3</sup>, with the lower values attributed to lower seasonal rainfall, but none considered to be in poor condition (Loehr et al. 2007). Similarly, body condition indices for adult Western Pond Turtles (*Actinemys marmorata*) ranged from 1.07–1.09 g/cm<sup>3</sup> in high quality streams but



1.01–1.05 g/cm<sup>3</sup> in lower quality streams (Ashton et al. 2015).

Based on the index developed specifically for desert tortoises (Nagy et al. 2002), mean body condition for our three treatment groups ranged from 0.554–0.586 g/cm<sup>3</sup>. Although none of our mean body conditions fell within values considered to reflect prime body condition in wild desert tortoises (0.60–0.70 g/cm<sup>3</sup>), they were well above values for dehydrated wild animals (0.40 g/cm<sup>3</sup>; Nagy et al. 2002). The direct-release animals in our study exhibited the greatest range of values (0.48–0.64 g/cm<sup>3</sup>), likely reflecting variation in availability or distribution of resources (shelter, forage, water) in the wild. It is worth noting that we were only able to obtain body condition measurements of DR animals that survived the first 6 months post-release, so we do not know to what extent body condition contributed to mortality of DR animals. In contrast, the indoor and outdoor head-starts exhibited much less variation in body condition (0.51–0.60 g/cm<sup>3</sup> and 0.54–0.63 g/cm<sup>3</sup>, respectively). Both indoor and outdoor-reared tortoises had frequent access to drinking water, which likely contributed to the more consistent body condition of animals from those treatments.

We found that smaller 2015 indoor-reared tortoises weighed less than outdoor-reared tortoises of similar size from the 2011 and 2012 cohorts; however, this difference diminished as tortoises in all groups approached 95 mm in MCL. The differences between treatments among smaller animals are likely due to differences in bone mass and shell ossification between the age groups, making the younger indoor head-starts lighter with respect to their volume compared to older outdoor head-starts (Arendt and Wilson 2000). Loehr et al. (2007) also noted that juvenile shells of *H. signatus* are less well-ossified (i.e., softer) than adults, giving juveniles a lower body mass to volume ratio than adults, which are similar in size to juvenile desert tortoises. In contrast, younger indoor head-starts that have attained 95 mm MCL in our study show no evidence of deficiency in volume-corrected mass compared to older outdoor-reared tortoises of the same size.

**Management implications.**—Indoor head-starting was successful in reducing the time required for desert tortoises to reach sizes approaching published recommendations for release (Nagy et al. 2015b; Hazard et al. 2015). Future studies that evaluate indoor-rearing for longer periods and/or incorporate a combination of indoor and outdoor husbandry (e.g., an entire year of indoor-rearing followed by a final year of outdoor rearing) may reveal the conditions needed to increase shell hardness among head-started tortoises while retaining the growth-promoting benefit of indoor-rearing in the first year of life. Furthermore, use of short-wave ultraviolet (UVB) lights coupled with

calcium and vitamin D<sub>3</sub> supplementation during indoor head-starting may promote better bone development and shell hardness in future efforts.

Because rapid growth may come with fitness costs (Jackson et al. 1976, Olsson and Shine, 2002), future efforts are likely to be most successful if they consider additional morphological metrics rather than size alone. Body condition and shell hardness indices are valuable metrics for evaluating head-starting efforts. Future head-start studies that monitor body condition can provide information to better evaluate the relationship between body condition and shell hardness, as body condition may increase as shells ossify. Both metrics will be important for helping to assess survival and growth in the wild and in linking release strategies with environmental conditions (i.e., seasonal/annual rainfall) and thus fine-tuning head-starting to benefit species recovery. Ultimately, the success of indoor rearing and head-starting in general can best be evaluated through post-release monitoring of survival and, eventually, reproduction.

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### LITERATURE CITED

- Abella, S.R., and K.H. Berry. 2016. Enhancing and restoring habitat for the Desert Tortoise *Gopherus agassizii*. *Journal of Fish and Wildlife Management* 7:1–25.
- Allison, L. 2012. Range-wide monitoring of the Mojave Desert Tortoise (*Gopherus agassizii*): 2008 and 2009

- reporting. Desert Tortoise Recovery Office, U.S. Fish and Wildlife Service, Reno, Nevada. 39 p.
- Arendt, J.D., and D.S. Wilson. 2000. Population differences in the onset of cranial ossification in Pumpkinseed (*Lepomis gibbosus*), a potential cost of rapid growth. *Canadian Journal of Fisheries and Aquatic Sciences* 57:351–356.
- Ashton, D.T., J.B. Bettaso, and H.H. Welch, Jr. 2015. Changes across a decade in size, growth, and body condition of Western Pond Turtle (*Actinemys marmorata*) populations on free-flowing and regulated forks of the Trinity River in Northwest California. *Copeia* 103:621–633.
- Beatley, J.C. 1974. Phenological events and their environmental triggers in Mojave Desert ecosystems. *Ecology* 55:856–863.
- Berry, K. 1986. Desert Tortoise (*Gopherus agassizii*) research in California, 1976–1985. *Herpetologica* 42:62–67.
- Boarman, W.I. 2003. Managing a subsidized predator population: reducing Common Raven predation on Desert Tortoises. *Environmental Management* 32:205–217.
- Burke, L.R. 2015. Head-starting turtles: learning from experience. *Herpetological Conservation and Biology* 10:299–308.
- Buhlmann, K.A., S.L. Koch, B.O. Butler, T.D. Tuberville, V.J. Palermo, B.A. Bastarache, and Z.A. Cava. 2015. Reintroduction and head-starting: tools for Blanding's Turtle (*Emydoidea blandingii*) conservation. *Herpetological Conservation and Biology* 10:436–454.
- Cagle, F.R. 1939. A system of marking turtles for future identification. *Copeia* 1939:170–173.
- Caillouet Jr., C.W., D.J. Shaver, and A.M. Landry, Jr. 2015. Kemp's Ridley Sea Turtle (*Lepidochelys kempii*) head-start and reintroduction to Padre Island National Seashore, Texas. *Herpetological Conservation and Biology* 10:309–377.
- Cohn, P. 1999. Saving the California Condor: years of effort are paying off in renewed hope for the species' survival. *Bioscience* 49:864–868.
- Curtin, A.J., G.R. Zug, and J.R. Spotila. 2009. Longevity and growth strategies of the Desert Tortoise (*Gopherus agassizii*) in two American deserts. *Journal of Arid Environments* 73:463–471.
- Daly, J.A. 2017. Indoor-rearing as a component of head-starting the Mojave Desert Tortoise (*Gopherus agassizii*). M.Sc. Thesis, University of Georgia, Athens, Georgia, USA. 97 p.
- Esque, T., K. Nussear, K. Drake, A. Walde, K. Berry, R. Averill-Murray, A. Woodman, W. Boarman, P. Medica, J. Mack, and J. Heaton. 2010. Effects of subsidized predators, resource variability, and human population density on Desert Tortoise populations in the Mojave Desert, USA. *Endangered Species Research* 12:167–177.
- Garcia-Berthou, E. 2001. On the misuse of residuals in ecology: testing regression residuals vs. the analysis of covariance. *Journal of Animal Ecology* 70:108–111.
- Gibbons, J.W. 1987. Why do turtles live so long? *Bioscience* 37:262–269.
- Gibbs, J.P., E.A. Hunter, K.T. Shoemaker, W.H. Tapia, and L.J. Cayot. 2014. Demographic outcomes and ecosystem implications of Giant Tortoise reintroductions to Espanola Island, Galapagos. *PLoS ONE* 9(10):1–15. <https://doi.org/10.1371/journal.pone.0110742>.
- Green, J.M. 2015. Effectiveness of head-starting as a management tool for establishing a viable population of Blanding's Turtles. M.Sc. Thesis, University of Georgia, Athens, Georgia, USA. 110 p.
- Griffiths, C.J., C.J. Jones, D.M. Hansen, M. Puttoo, R.V. Tatayah, C.B. Muller, and S. Harris. 2010. The use of extant non-indigenous tortoises as a restoration tool to replace extinct ecosystem engineers. *Restoration Ecology* 18:1–7.
- Hazard, L., and D. Morafka. 2002. Comparative dispersion of neonate and headstarted juvenile Desert Tortoises (*Gopherus agassizii*): a preliminary assessment of age effects. *Chelonian Conservation and Biology* 44:135–147.
- Hazard, L.C., D.J. Morafka, and L.S. Hillard. 2015. Post-release dispersal and predation of head-started juvenile Desert Tortoises (*Gopherus agassizii*): effect of release site distance on homing behavior. *Herpetological Conservation and Biology* 10:504–515.
- Heppl, S., L. Crowder, and D. Crouse. 1996. Models to evaluate headstarting as a management tool for long-lived turtles. *Ecological Applications* 6:556–565.
- Jackson, C.G., J.A. Trotter, T.H. Trotter, and M.W. Trotter. 1976. Accelerated growth rate and early maturity in *Gopherus agassizii* (Reptilia: Testudines). *Herpetologica* 32:139–145.
- Jarchow, J.L., H.E. Lawler, T.R. Van Devender, and C.S. Ivanyi. 2002. Care and diet of captive Sonoran Desert Tortoises. Pages 289–311 *In* The Sonoran Desert Tortoise. Van Devender, T.R. (Ed.). University of Arizona Press, Tucson, Arizona, USA.
- Jarvie, S., A.M. Senior, S.C. Adolph, P.J. Seddon, and A. Cree. 2015. Captive rearing affects growth but not survival in translocated juvenile Tuatara. *Journal of Zoology* 297:184–193.
- Jennings, W.B., and K.H. Berry. 2015. Desert Tortoises (*Gopherus agassizii*) are selective herbivores that track the flowering phenology of their preferred food plants. *PLoS ONE* 10:1–32. <https://doi.org/10.1371/journal.pone.0116716>.

- Lannoo, M.J. (Ed.). 2005. Amphibian Declines: The Conservation Status of United States Species. University of California Press, Berkeley, California, USA.
- Lindenmayer, D., and B. Scheele. 2017. Do not publish: limiting open access information on rare and endangered species will help to protect them. *Science* 356:800–801.
- Loehr, V., B. Henen, and M. Hofmeyr. 2004. Reproduction of the smallest tortoise, the Namaqualand Speckled Padloper, *Homopus signatus signatus*. *Herpetologica* 60:759–761.
- Loehr, V.J.T., M.D. Hofmeyr, and B.T. Henen. 2007. Annual variation in the body condition of a small, arid-zone tortoise, *Homopus signatus signatus*. *Journal of Arid Environments* 71:337–349.
- Nafus, M.G., B.D. Todd, K.A. Buhlmann, and T.D. Tuberville. 2015. Consequences of maternal effects on offspring size, growth and survival in the Desert Tortoise. *Journal of Zoology* 297:108–114.
- Nafus, M.G., T.D. Tuberville, K.A. Buhlmann, and B.D. Todd. 2013. Relative abundance and demographic structure of Agassiz's Desert Tortoise (*Gopherus agassizii*) along roads of varying size and traffic volume. *Biological Conservation* 162:100–106.
- Nafus, M.G., T.D. Tuberville, K.A. Buhlmann, and B.D. Todd. 2017. Precipitation quantity and timing affect native plant production and growth of a key herbivore, the Desert Tortoise, in the Mojave Desert. *Climate Change Responses* 4:1–10.
- Nagy, K., B. Henen, B. Devesh, and I. Wallis. 2002. A condition index for the Desert Tortoise (*Gopherus agassizii*). *Chelonian Conservation and Biology* 4:425–429.
- Nagy, K.A., S. Hillard, S. Dickson, and D.J. Morafka. 2015a. Effects of artificial rain on survivorship, body condition, and growth of head-started Desert Tortoises (*Gopherus agassizii*) released to the open desert. *Herpetological Conservation and Biology* 10:535–549.
- Nagy, K.A., L.S. Hillard, M.W. Tuma, and D.J. Morafka. 2015b. Head-started Desert Tortoises (*Gopherus agassizii*): movements, survivorship and mortality causes following their release. *Herpetological Conservation and Biology* 10:203–215.
- Nagy, K., M. Tuma, and L. Hillard. 2011. Shell hardness measurements in juvenile Desert Tortoises *Gopherus agassizii*. *Herpetological Review* 42:191–195.
- Nussear, K.E., T.C. Esque, R.D. Inman, L. Gass, K.A. Thomas, C.S.A. Wallace, J.B. Blainey, D.M. Miller, and R.H. Webb. 2009. Modeling habitat of the Desert Tortoise (*Gopherus agassizii*) in the Mojave and parts of the Sonoran Deserts of California, Nevada, Utah, and Arizona. U.S. Geological Survey Open-File Report 2009-1102. Sacramento, California, USA. 18 p.
- Olsson, M., and R. Shine. 2002. Growth to death in lizards. *Evolution* 56:1867–1870.
- Peaden, J.M., T.D. Tuberville, K.A. Buhlmann, M.G. Nafus, and B.D. Todd. 2015. Delimiting road-effect zones for threatened species: implications for mitigation fencing. *Wildlife Research* 42:650–659.
- Pollock, K.H., S.R. Winterstein, C.M. Bunck, and P.D. Curtis. 1989. Survival analysis in telemetry studies: the staggered entry design. *Journal of Wildlife Management* 53:7–15.
- Quinn, D.P., K.A. Buhlmann, J.B. Jensen, T.M. Norton, and T.D. Tuberville. 2018. Post-release movement and survivorship of head-started Gopher Tortoises. *Journal of Wildlife Management* 82:1545–1554.
- R Core Team. 2014. R: a language and environment for statistical computing. Foundation for Statistical Computing, Vienna, Austria. <http://www.R-project.org>.
- Reed, J., N. Fefferman, and R. Averill-Murray. 2009. Vital rate sensitivity analysis as a tool for assessing management actions for the Desert Tortoise. *Biological Conservation* 142:2710–2717.
- Shine, R., M.P. Lemaster, I.T. Moore, M.M. Olsson, and R.T. Mason. 2001. Bumpus in the snake den: effects of sex, size, and body condition on mortality of Red-sided Garter Snakes. *Evolution* 55:598–604.
- Sinn, D.L., L. Cawthen, S.M. Jones, C. Pukk, and M.E. Jones. 2014. Boldness towards novelty and translocation success in captive-raised, orphaned Tasmanian Devils. *Zoo Biology* 33:36–48.
- Spencer, R.J., J.U. Van Dyke, and M.B. Thompson. 2017. Critically evaluating best management practices for preventing freshwater turtle extinctions. *Conservation Biology* 31:1340–1349.
- Todd, B.D., B. Halstead, L.P. Chiquoine, J.M. Peaden, K.A. Buhlmann, T.D. Tuberville, and M.G. Nafus. 2016. Habitat selection by juvenile Mojave Desert Tortoises. *Journal of Wildlife Management* 80:720–728.
- Tuberville, T.D., T.M. Norton, K.A. Buhlmann, and V. Greco. 2015. Head-starting as a management component for Gopher Tortoises (*Gopherus polyphemus*). *Herpetological Conservation and Biology* 10:455–471.
- Turner, F., P. Medica, and R. Bury. 1987. Age-size relationships of Desert Tortoises (*Gopherus agassizii*) in southern Nevada. *Copeia* 1987:974–979.
- Turner, F., P. Medica, and C. Lyons. 1984. Reproduction and survival of the Desert Tortoise (*Scaptochelys agassizii*) in Ivanpah Valley, California. *Copeia* 1984:811–820.
- U.S. Fish and Wildlife Service (USFWS). 1990. Endangered wildlife and plants: determination of

Daly et al.—Growth and condition of head-started Desert Tortoises.

threatened status for the Mojave population of the Desert Tortoise. Federal Register 55:12178–12191. U.S. Fish and Wildlife Service (USFWS). 2008. Draft revised recovery plan for the Mojave population of the Desert Tortoise. U.S. Fish and Wildlife Service, Sacramento, California, USA.

U.S. Fish and Wildlife Service (USFWS). 2011. Revised recovery plan for the Mojave population of the Desert Tortoise. U.S. Fish and Wildlife Service, Sacramento, California, USA.

Vander Haegen, W.M., S.L. Clark, K.M. Perillo, D.P. Anderson, and H.L. Allen. 2009. Survival and causes of mortality of head-started Western Pond Turtles on Pierce National Wildlife Refuge, Washington. *Journal of Wildlife Management* 73:1402–1406.

Vikash, T., N. Zuël, N.C. Cole, C. Griffiths, and C.G. Jones. 2018. Introduction to Ile aux Aigrettes, Mauritius, of the Aldabra Giant Tortoise as an ecological replacement for the extinct Mauritian Tortoise. Pp. 87–91 *In* Global Re-introduction Perspectives: 2018. Case-Studies from Around the Globe. Soorae, P.S. (Ed.). International Union for Conservation of Nature, Gland, Switzerland.

Wiesner, C.S., and C. Iben. 2003. Influence of environmental humidity and dietary protein on pyramidal growth of carapaces in African Spurred Tortoises (*Geochelone sulcata*). *Journal of Animal Physiology and Animal Nutrition* 87:66–74.



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